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1 **The enrichment of an alkaliphilic biofilm consortia capable of the anaerobic**
2 **degradation of isosaccharinic acid from cellulosic materials incubated within an**
3 **anthropogenic, hyperalkaline environment.**

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11 Running title = Isolation of an alkaliphilic biofilm consortia.

12 **Abstract**

13 Anthropogenic hyper-alkaline sites provide an environment that is analogous to
14 proposed cementitious geological disposal facilities (GDF) for radioactive waste. Under
15 anoxic, alkaline conditions cellulosic wastes will hydrolyse to a range of cellulose
16 degradation products (CDP) dominated by isosaccharinic acids (ISA). In order to investigate
17 the potential for microbial activity in a cementitious GDF, cellulose samples were incubated
18 in the alkaline (~pH 12), anaerobic zone of a lime kiln waste site. Following retrieval, these
19 samples had undergone partial alkaline hydrolysis and were colonised by a Clostridia
20 dominated biofilm community, where hydrogenotrophic, alkaliphilic methanogens were also
21 present. When these samples were used to establish an alkaline CDP fed microcosm, the
22 community shifted away from Clostridia, methanogens became undetectable and a flocculate

23 community dominated by *Alishewanella sp.* established. These flocs were composed of
24 bacteria embedded in polysaccharides and protein stabilised by extracellular DNA. This
25 community was able to degrade all forms of ISA with >60% of the carbon flow being
26 channelled into extracellular polymeric substance (EPS) production. This study demonstrated
27 that alkaliphilic microbial communities can degrade the CDP associated with some
28 radioactive waste disposal concepts at pH 11. These communities divert significant amounts
29 of degradable carbon to EPS formation, suggesting that EPS has a central role in the
30 protection of these communities from hyper-alkaline conditions.

31

31

32 **Introduction**

33 The UK's national nuclear waste legacy contains approximately 290 000 m³ (N.D.A.,
34 2013) of intermediate level radioactive wastes (ILW) which includes an estimated (~2000
35 tonnes) (N.D.A., 2010a) of cellulosic materials (wood, paper and cloth) (Humphreys et al.,
36 2010a). One of the proposed strategies for the disposal of this ILW is a deep geological
37 disposal facility (GDF) (N.D.A., 2010a) employing a multi-barrier system which is likely to
38 include a cement based backfill (Chapman and Hooper, 2012). Upon the closure of such a
39 facility, groundwater ingress combined with corrosion processes will result in the
40 development of a chemically reducing high pH (pH 12.5) environment (Libert et al., 2011,
41 N.D.A., 2010b). Under these conditions the cellulose portion of ILW is expected to undergo
42 chemical, alkaline hydrolysis to form a variety of cellulose degradation products (CDP)
43 (Knill and Kennedy, 2003, Humphreys et al., 2010a).

44 CDP are comprised of the alpha and beta diastereomers of isosaccharinic acid (ISA),
45 alongside other small chain organic compounds including acetic acid (Van Loon and Glaus,
46 1997, Motellier et al., 1998, Knill and Kennedy, 2003). The diastereomers of ISA are of
47 significance when considering the performance of a GDF as they possess the ability to
48 enhance the mobility of a range of radionuclides, including nickel, thorium, plutonium and
49 uranium through complexation (Greenfield et al., 1991, Allard and Ekberg, 2006, Warwick et
50 al., 2003). In addition, the hemicellulose fraction of cellulosic waste components will also
51 undergo anoxic, alkaline hydrolysis to form an additional 5-carbon form of ISA, known as
52 xyloisosaccharinic acid (X-ISA) (Almond et al., 2012). Recent work by Randall et al (2013)
53 suggests that X-ISA does not have the same complexation properties as the alpha and beta

54 forms of ISA but could, however, represent a source of organic carbon available for microbial
55 metabolism.

56 Although the harsh geochemical conditions of an ILW-GDF place limitations upon
57 microbial life it may not prevent microbes from colonising a facility. An investigation of an
58 anthropogenic analogue of an ILW-GDF at a hyperalkaline contaminated site in Buxton, UK
59 where ISA is generated in-situ (Rout et al., 2015) has revealed a microbially active site
60 despite porewaters of up to pH 13 (Burke et al., 2012). The range of microbes present within
61 the background sediments is diverse, with organisms within the Phyla Bacteroidetes,
62 Proteobacteria and Firmicutes consistently making up large proportions of the sediment
63 taxonomic profiles (Burke et al., 2012, Williamson et al., 2013, Bassil et al., 2014). The
64 subsequent culturing of these sediments has shown that these communities are able to utilise
65 the alpha form of ISA as a substrate under aerobic, nitrate, iron reducing (Bassil et al., 2014)
66 and methanogenic conditions (Rout et al., 2015). Sulphate reduction appears to be inhibited at
67 pH >10 (Bassil et al., 2014), however, the utilisation of ISA under sulphate reducing
68 conditions has been observed at neutral pH indicating that this limitation is thermodynamic
69 (Rizoulis et al., 2012, Rout et al., 2014). The heterogeneity of ILW and its compaction in
70 grout may limit the availability of higher energy terminal electron acceptors such as nitrate
71 and ferric iron, with the inundating ground water also depleted in these electron acceptors due
72 to its passage through the microbial thermodynamic ladder (Bethke et al., 2011).
73 Fermentation processes and subsequent methanogenesis therefore represent the most likely
74 conditions to dominate an ILW-GDF

75 Microbes in nature can be found in biofilms of mixed syntrophic communities, with
76 microbial biofilms found in a diverse range of environments (Summons et al., 2015, Urbieta
77 et al., 2015). The secretion of extracellular polymeric substance (EPS) such as
78 polysaccharides, proteins, lipids and nucleic acids during biofilm formation assist in bacterial

79 survival and propagation (Flemming and Wingender, 2010) and confer an increased
80 resistance to environmental stresses such as pH and temperature fluctuations, desiccation and
81 UV radiation (Ordoñez et al., 2009, Gorlenko et al., 2004, Rodrigues et al., 2006, Jones et al.,
82 1994, Conrad et al., 2014). When considering the colonisation of an ILW-GDF, the ability of
83 microbes to migrate and adhere to niche areas such as ungrouted surfaces may allow for both
84 microbial survival and growth under extreme alkaline conditions (Humphreys et al., 2010b).
85 The aim of this work was to culture, *in situ*, a biofilm forming consortium capable of
86 colonising cellulosic materials under anoxic, hyper-alkaline conditions and to determine its
87 ability to degrade CDP, which represent the primary organic carbon source within an ILW-
88 GDF.

89 **Methods**

90 Cellulose cotton preparation

91 In order to prepare the cellulose cotton for incubation, raw woven cotton fabric
92 (Greige) was treated with NaOH to saponify the natural waxes along with an alkali stable
93 phosphate ester detergent to emulsify the suspended impurities. Further treatment with NaOH
94 and phosphonate stabilised H₂O₂ was carried out to bleach the fabric. The cotton was then
95 rinsed, neutralised under acetic acid before finally being rinsed, dried and autoclaved at
96 121°C prior to use.

97 Analogue site investigation

98 During May 2014 a 2.2cm Ø borehole was hand drilled to an approximate depth of
99 0.5m into an area inundated with alkaline leachate at Brook Bottom, Harpur Hill, Buxton, UK
100 (Figure 1). An inert plastic liner with a perforated lower section was placed into the borehole.
101 Approximately 5g of sterile treated cellulose cotton was loaded into a nylon mesh bag and

102 placed at the bottom of the borehole. After a period of 3 months the cotton was recovered
103 along with sediment and porewater samples from the immediate vicinity of the sample. *In*
104 *situ* pH and Eh values were determined prior to sample recovery using a handheld portable
105 pH meter with calibrated electrodes and an InLab Redox Micro probe (Mettler Toledo, UK)
106 tested in accordance with BS ISO 11271:2002 (B.S.I, 2002). All recovered materials were
107 sealed in airtight containers along with anaerobic gas packs (Anaerogen, Oxoid, UK) for
108 transport. Sediment and porewater samples were stored at -20°C until analysis and cotton not
109 used for immediate studies was stored at -20°C in a solution of 140mL of ultrapure water,
110 10mL of 1M TRIS-HCl (pH7.5) and 250ml of 96% ethanol after an overnight fixation step in
111 4% paraformaldehyde in phosphate buffered saline.

112 Porewater, cotton and sediment ISA content was determined as previously described
113 by Rout et al (2015, 2014) against ISA standards in the alpha, beta and xylo conformations
114 (Almond et al., 2012, Shaw et al., 2012). C1-8 volatile fatty acid (VFA) content of both the
115 sediment and cotton was determined using a standard extraction method outlined in Eaton et
116 al (2005) and analysed via GC-FID as described by Rout et al (2014).

117 Microscopy

118 Scanning electron microscopy was undertaken using a JEOL JSM-6060LV
119 microscope (JEOL, USA). Samples were dehydrated using a serial ethanol dilution of 25 %,
120 50 %, 75% and 100 % for 2 minutes per step then sputter coated via a gold palladium plasma
121 (CA7625 Polaron, Quorum Technologies Ltd, UK). Fluorescence microscopy was carried out
122 using an Olympus BX41 laboratory microscope (Olympus, USA). Live dead staining was
123 carried out using the BAC light Live/dead kit (Life technologies, UK), fluorescein
124 isothiocyanate (FITC) (Sigma-Aldrich, UK) staining was used for protein and visualisation of
125 individual bacteria cells and the polysaccharide components was achieved using ethidium

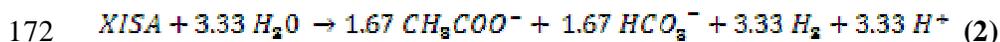
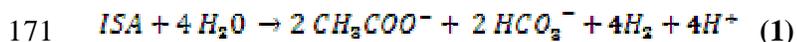
126 bromide and Calcofluor White (Sigma-Aldrich, UK) staining, respectively. For DNase
127 digestion, microcosm fluid (1mL) was centrifuged at 10,000 xg for 1 minute and re-
128 suspended in ultrapure water (1mL). A 10 fold dilution of this was then subjected to digestion
129 by DNase using a DNase 1 kit (Sigma-Aldrich, UK).

130 Microcosm

131 In order to investigate ISA degradation, approximately 1g of colonised cotton was
132 washed with 10mL N₂ purged sterile PBS under an inert environment to remove any transient
133 microorganisms. The washed cotton was then added to a continuously stirred microcosm
134 containing 175mL of pre-reduced 10% CDP and 90% mineral media (B.S.I, 2005) at pH 11
135 and 20°C that had been purged with nitrogen and maintained with a nitrogen headspace to
136 ensure anoxic conditions. CDP was produced as previously described by Rout et al (2014).
137 The microcosm was brought up to a final volume of 250mL by feeding 25mL of CDP every 2
138 weeks with the pH adjusted using 4M NaOH every 7 days. After this period the cotton was
139 removed and the microcosm was switched to a 10% waste/feed cycle with CDP every 2
140 weeks. The microcosm was maintained with a nitrogen atmosphere and all reagents were
141 reduced prior to use with disodium sulfide nonahydrate (Sigma-Aldrich, UK) and sodium
142 dithionite (Fisher, UK) as per BS ISO 14853:2005 (B.S.I, 2005) and stored under nitrogen.
143 Resazurin redox indicator (Fisher, UK) present within the mineral media provided an
144 indication of anaerobic conditions within the microcosm and all manipulations of the
145 microcosm were carried out under a stream of nitrogen to maintain anoxic conditions.
146 Sufficient time (50 weeks) was allowed for the microcosm chemistry to stabilise and also to
147 allow for the washout of any transient microorganisms. The microcosm was sampled every 2
148 days over 2 feed/waste cycles to determine the ISA and VFA content. For each sample period
149 microcosm fluid (1mL) was taken, centrifuged at 10,000 xg for 1 minute and the supernatant
150 filter sterilised using a 0.45µm syringe filter (Sartorius, UK) and stored at -20°C prior to

151 analysis. The gas headspace (75ml starting volume) was sampled every 2 days with the
152 composition determined via gas chromatography using Agilent 6850 gas chromatograph
153 (Hewlett Packard, UK) fitted with a HP-Plot/Q+ PT column and thermal conductivity
154 detection (TCD). Headspace gas (100µL) was removed using a lockable gas syringe from the
155 microcosm and passed through the column under the following conditions: initial temperature
156 of 60°C for 2 minutes, followed by an increase to 120°C at a ramp rate of 30°C min⁻¹ with a
157 detector temperature of 250°C. Gas headspace pressure was measured using a digital
158 manometer (TPI, UK) before gas sample periods.

159 Microcosm fluid (1mL) containing the suspended flocs was taken on days 0, 7 and 14
160 and spun at 10,000 xg for 1 minute for ATP/biomass detection using a 3M™ Clean-Trace™
161 Biomass Detection Kit and Luminometer employing a modified method (3M, UK). The pellet
162 was washed once with pH 4 PBS and reconstituted in pH 7 PBS to remove interference from
163 excess alkalinity and salts. Following analysis, CFU mL⁻¹ and dry weight biomass (DW)
164 were calculated against a standard curve of *E.coli* K12 concentrations. In addition, a set of
165 control microcosms amended with 50µg mL⁻¹ chloramphenicol were prepared and were
166 sampled as per the test microcosms. The controls served as an abiotic comparison for the
167 elimination of sorption and precipitation events. All data was processed in Microsoft Excel
168 with calculated means and associated standard error shown in all relevant results. Carbon
169 flow calculations were undertaken using balanced equations 1 and 2 for the fermentation of
170 ISA to acetate and hydrogen.



173 Preparation of 16S rDNA clone libraries

174 Total genomic DNA was extracted from the cotton and microcosm using a Powersoil
175 DNA extraction kit (Mo-BIO, Carlsbad, US) with the following modifications. For the cotton
176 approximately 0.25g was washed with pH 7 PBS and loaded into a glass bead tube with
177 100 μ L β - mercaptoethanol and the bead beating step extended to 1 hour in order to overcome
178 dampening effects introduced by the material. For genomic DNA extraction from the
179 microcosm, 25mL of fluid was centrifuged at 5000 xg for 15min and the pellet re-suspended
180 in 25mL pH 4 PBS. The sample was then centrifuged again at 5000 xg for 15 minutes and re-
181 suspended in 2mL of pH 7 PBS. 1mL of the concentrated sample was transferred to a 1.5mL
182 tube and centrifuged again at 10,000 xg for 1 minute, after which the supernatant was
183 removed and the cell pellet re-suspended in the reaction fluid provided in the glass bead tubes
184 of the Powersoil kit. The resulting mixture was then transferred back to a glass bead tube and
185 bead beaten with 100 μ L β -mercaptoethanol for an increased time of 20 minutes to overcome
186 clogging due to the EPS and then run as per the supplier's instruction. These modifications
187 were found to increase the yield and purity of DNA obtained from both samples by removing
188 excess salts, inhibiting nucleases and neutralising the samples.

189 Purified genomic DNA was quantified and quality checked by spectroscopic methods
190 and used as a template to amplify the 16s rRNA gene. A ~1500bp fragment of the Eubacterial
191 16S rRNA gene was amplified using broad specificity primers pA and pH (Edwards et al.,
192 1989) and a ~750bp fragment of the archaeal 16S rRNA gene was amplified using primers Ar
193 and Af (Gantner et al., 2011). PCR reactions were carried out using BIOMIX red master mix
194 (BIOLINE, UK) with PCR fragments purified via a Qiaquick PCR purification kit (Qiagen,
195 UK) and visualised using a 1.0% agarose TAE gel with SYBR® Safe staining (Life
196 technologies, UK). PCR products were ligated into the standard cloning vector PGEM-T easy
197 (Promega, US) and transformed into *E.coli* JM109 competent cells (Promega, US).
198 Transformed cells were grown on Luria Bertani (LB) agar containing 100 μ g mL⁻¹ ampicillin

199 overlaid with 40 μ L of 100mM IPTG and 40 μ L of 40mg mL⁻¹ X-GAL (5-bromo-4-chloro-3-
200 indolyl- β -D-galactopyranoside) in N’N dimethylformamide for blue-white colour screening.
201 Insert containing colonies were transferred to 96 well plates containing LB agar with 150mg
202 mL⁻¹ ampicillin and sequenced using Sanger sequencing technology (GATC Biotech,
203 Germany). Inserts were amplified using a T7 forward primer and the resulting 16S rRNA
204 gene sequences aligned using the multiple sequence alignment package MUSCLE
205 (www.ebi.ac.uk/Tools/services/web/toolform.ebi?tool=muscle) and chimera checked using
206 the UCHIME component of the Mothur suite, where chimeric sequences were omitted from
207 the analysis (Schloss et al., 2009). Sequences were analysed against the NCBI database using
208 Basic Local Alignment Search Tool (MegaBLAST) utilising the 16S ribosomal RNA
209 sequences for Bacteria and Archaea (Altschul et al., 1997). Phylogenetic families were then
210 determined at a 95% confidence level by comparison with the Ribosomal Database Project
211 (Cole et al., 2009).

212 Nucleotide accession numbers

213 The 16S rRNA sequence data from the colonised cotton has been submitted to
214 GenBank under accession numbers KP263977 - KP264111 and the microcosm sequences
215 under the numbers KP728118 - KP728176.

216 **Results**

217 Chemical and physiological cotton analysis

218 The pH in the vicinity of the cotton samples was between pH 11.5 and 12 and redox
219 measurements were found to be negative in both the associated sediment (-77mV) and
220 porewaters (-66mV). Both the alpha and beta forms of ISA were extracted from the cotton
221 (>0.5mg (g dry wt)⁻¹), the sediment (>0.5mg (g dry wt)⁻¹)) and porewater (7.64mg L⁻¹ alpha,

222 6.82mg L⁻¹ beta) (Table 1) indicating in-situ alkaline cellulose hydrolysis (Knill and
223 Kennedy, 2003).

224 The surfaces of the colonised cotton showed areas of EPS indicative of biofilm
225 formation and surface associated mineral precipitates (Figure 2B and Figure S1) with
226 individual viable bacterial cells being visible on some fibres (Figure 2C and D).

227 16S rDNA profile of colonised cotton

228 The cotton's Eubacterial clone library (Table S1) was dominated by the Order
229 Clostridiales which represented 58% of the clones obtained (n=67, Figure 3A). Of these
230 Clostridia, 33 sequences most closely matched organisms from the family Clostridiaceae 2,
231 where 13 sequences most closely matched *Clostridium formicaceticum* strain DSM 92 (95%
232 sequence similarity) and a further 10 to *Anaerovirgula multivorans* strain SCA (97%
233 sequence similarity). The remaining 9 clones most closely matched sequences belonging to
234 the genus *Alkaliphilus*, of which 8 were closely related to *Alkaliphilus oremlandii* strain
235 OhILAs (91-93% sequence similarity) and 1 related to *Alkaliphilus transvaalensis* strain
236 SAGM1 (98% sequence similarity). The remaining clones of the Clostridia were represented
237 by sequences most closely related to organisms from the families *Clostridium insertae sedis*
238 *XI* (3 sequences) and *Clostridium insertae sedis XIV* (2 sequences).

239 The remainder of the clone library was made up of a diverse range of taxonomic
240 families (Table S1) including: representatives from Brucellaceae, primarily related to
241 *Ochrobactrum anthropi* strain ATCC 49188 (96-99% sequence similarity);
242 Corynebacteriaceae, dominated by *Corynebacterium marinum* strain D7015 (98-99%
243 sequence match similarity); and the Bacillaceae 1, dominated by *Bacillus pseudofirmus* strain
244 OF4 (89-99% sequence similarity).

245 The Archaeal clone library (Table S2) was dominated (93%) by sequences most
246 closely matching *Methanobacterium alcaliphilum* strain NBRC 105226 (99% sequence
247 similarity) (n=68, Figure 3B). The remaining sequences were most closely related to
248 *Methanomassiliicoccus luminyensis* strain B10 (4 sequences 89% sequence match) and
249 *Methanosarcina mazei* Go1 (99% sequence similarity).

250 CDP driven microcosms

251 The microcosm demonstrated significant degradation of ISA at pH 11.0 over 2
252 waste/feed cycles (Figure 4) with first order rate constants of $3.33 \times 10^{-2} \text{ day}^{-1}$ ($\text{SE} \pm 2.0 \times 10^{-2}$)
253 2) for alpha, $9.36 \times 10^{-2} \text{ day}^{-1}$ ($\text{SD} \pm 2.2 \times 10^{-2}$) for beta and $6.78 \times 10^{-2} \text{ day}^{-1}$ ($\text{SE} \pm 2.85 \times 10^{-2}$)
254 for X-ISA. Acetate was the only VFA detected and gradually accumulated in the system
255 reaching a peak of 2.06 mmoles ($\text{SE} \pm 0.2$), similarly hydrogen gas accumulated in the
256 headspace over the course of the feed cycle reaching 1.00 mmoles ($\text{SE} \pm 0.04$). Neither
257 carbon dioxide nor methane was detected in the headspace of the microcosm, however,
258 soluble inorganic carbon increased within the system (data not shown) with the pH after each
259 cycle having an average pH of 10.80 ($\text{SE} \pm 0.4$). The CDP fed microcosm inoculated with the
260 colonised cotton was dominated by polymicrobial flocs with fluorescence microscopy
261 showing microbial cells embedded in an EPS composed of protein, polysaccharide and
262 extracellular DNA (eDNA) (Figure 5).

263 Measurement of the ATP concentration of the microcosm showed that cell density
264 increased over the feed/waste cycles (Table S4) indicating that a portion of the organic
265 carbon was used for the generation of both cell biomass and EPS. Carbon flow calculations
266 (Rittmann and Mccarty, 2001) based on the degradation of ISA showed 23.7% of the carbon
267 was converted to acetate and 12.1% converted to carbonate from energy generating
268 processes, 0.5% was converted to cell biomass and a further 63.7% was theorised to be

269 involved in processes relating to EPS production. The yield of dry cell biomass was 0.012 mg
270 (mg ISA)⁻¹ degraded, the system could not be stoichiometrically balanced due to the
271 unknown composition of the flocculate EPS material. Comparison of the samples amended
272 with chloramphenicol showed no ISA degradation and the production of acetate and
273 hydrogen was not detected (Figure S2) indicating that ISA degradation was via microbial
274 activity rather than chemical processes or sorption.

275 Microcosm clone library

276 The microcosm microbial populations demonstrated a significant shift away from that
277 associated with the emplaced cotton samples, with Archaeal taxa no longer being detectable
278 and the Eubacterial population no longer dominated by the Clostridiales. The environmental
279 and physiological constraints imposed within the microcosm resulted in a population
280 dominated by clones of *Alishewanella jeotgali* strain MS1 (99% sequence similarity) from
281 the family Alteromonadaceae (Table S3, Figure 6). The remaining clones included
282 representatives of the family Bacillaceae, most closely matching *Bacillus pseudofirmus* strain
283 OF4 (98% sequence similarity) and *Alkaliphilus crotonatoxidans* strain B11-2 (98% sequence
284 similarity) of the family Clostridiaceae 2.

285 **Discussion**

286 Previous authors noted the presence of an organic electron donor within the soils at
287 Harpur Hill that allowed for electron flow into nitrate and iron reducing processes at depth
288 (Burke et al., 2012). The generation of CDPs from the site's soil organic matter has been
289 demonstrated (Rout et al., 2015) and in this study the addition of cotton cellulose resulted in
290 its partial alkaline hydrolysis to CDPs with the concentration of alpha and beta ISA in the
291 porewater and sediments being higher than those measured by Rout et al (2015). This
292 supports the concept that the hyper-alkaline conditions created at this site are capable of

293 generating CDP. The presence of acetate, a common end product of ISA fermentation (Rout
294 et al., 2015, Bassil et al., 2014, Rout et al., 2014), in the porewater, sediment and cotton
295 indicated an active anaerobic microbial community in the immediate proximity of the cotton
296 even though the ambient pH was between pH 11.5 and 12.

297 Cotton fibres were covered with large areas of EPS indicative of biofilm formation
298 (Figure 2A and B) with individual cells being only rarely visible (Figure 2C). This is a
299 marked contrast to the colonisation of cotton incubated in a landfill site under neutral
300 anaerobic conditions reported by McDonald et al (2012) where fibres were heavily colonised
301 with cells and exhibited the characteristic pits and grooves associated with microbial
302 cellulose hydrolysis. The reduced colonisation of the cotton under the hyperalkaline
303 conditions present at the site are further illustrated by the live/dead staining of the cotton
304 (Figure 2D) which revealed a low density of live cells on the individual cotton fibres and
305 within the surrounding biofilm material. Previous work by Grant et al (2002) demonstrated
306 the ability of alkaliphilic microorganisms to form a biofilm upon the surface of the
307 cementitious materials presumably to provide a degree of protection from the alkaline
308 stresses imposed by the local environment. This formation of EPS as a response to
309 hyperalkaline conditions is replicated in these microcosm studies where a polymicrobial,
310 eDNA stabilised floc based population developed (Figure 5A). The importance of EPS
311 generation in this system is illustrated by the fact that >60% of the available carbon is
312 diverted to EPS formation, a finding similar to the carbon distribution in biofilm systems
313 reported by Jahn and Nielsen (1998).

314 The microbial flocs were composed of an EPS containing protein, polysaccharides
315 and eDNA. Polysaccharides are a common component of EPS and moderate a range of
316 bacterial biofilm properties including adhesion, cell aggregation, cohesive nature, protection
317 as well as the sorption of organic compounds and inorganic ions (Flemming and Wingender,

318 2010). Imaging of the polysaccharide component revealed its distribution throughout the
319 flocculate with large globular like structures (Figure 5A). DNase treatment caused the loss of
320 these structures resulting a less compact structure of cells associated with polysaccharide,
321 indicating a relationship between the eDNA and the distribution of the polysaccharide
322 components (Figure 5B). The role of eDNA within biofilms appears to serve a number of
323 functions (Dominiak et al., 2011), in this case it is likely to aid the structure and function of
324 the flocculate community (Gloag et al., 2013). The presence of eDNA within the flocculate
325 structure is also likely to act as a phosphate store for the constituent microbial consortia
326 (Dell'anno and Danovaro, 2005). Calcium ions are abundant at the site and as such the
327 interaction between eDNA and these ions is likely to promote cell aggregation and biofilm
328 formation within these alkaliphilic cultures (Das et al., 2014). This is illustrated by the fact
329 that treatment of the flocs with DNase resulted in the loss of flocculate stability (Figure 5A
330 and B). Imaging of the protein component of the flocs showed large concentrated areas of
331 protein within the flocculate (Figure 5C). Protein serves a wide range of functions within
332 biofilm including the permitting of redox activity, protection from environmental conditions,
333 enzymatic reactions and sorption of organic compounds sorption and inorganic ions
334 (Flemming and Wingender, 2010).

335 The presence of the cotton cellulose within the sediments selected for organisms of
336 the Order Clostridia which contrasts with previous investigations of the background
337 sediments where a larger degree of taxonomic diversity was observed (Bassil et al., 2014,
338 Williamson et al., 2013), presumably due to greater diversity of energy sources and
339 colonisation from surrounding pasture land. Of the Clostridiaceae 2 species identified,
340 *Clostridium formicaceticum* has broad spectrum carbohydrate fermentation capabilities
341 (Andreesen et al., 1970), but was not previously associated with alkaline conditions. This
342 contrasts with species from the genera *Anaerovirgula* and *Alkaliphilus* which have all been

343 previously associated with alkaline sites (Fisher et al., 2008, Takai et al., 2001, Pikuta et al.,
344 2006).

345 The Archaeal population associated with the cotton was dominated by
346 hydrogenotrophic, alkaliphilic *Methanobacterium* sp. showing sequence similarity to
347 *Methanobacterium alcaliphilum* (Worakit et al., 1986). These findings are in agreement with
348 clone libraries generated from microcosms previously developed from sediment samples
349 from the same site (Rout et al., 2015). Although these organisms are able to utilise acetate as
350 a growth factor (Kotelnikova et al., 1998, Wu et al., 1992), they are incapable of acetoclastic
351 methanogenesis which accounts for the accumulation of acetic acid in extracts from the
352 cotton and surrounding sediment and porewaters. In addition a small number of sequences
353 showing similarity to *Methanomassiliicoccus luminyensis* (Dridi et al., 2012) and
354 *Methanosarcina* sp. (Maestrojuan et al., 1992) were also detected.

355 The microbial population established in the microcosm was much less diverse than
356 that present on the cotton samples with the almost complete removal of Clostridia and the
357 total loss of methanogens from the system. This resulting fermentative system was dominated
358 (95% of clones) by organisms most closely related to *Alishewanella* sp., which was a minor
359 component (3% of clones) of the population present on the colonised cotton. This facultative
360 anaerobic genus is most commonly associated with fermented seafood, but has also been
361 isolated from landfill soils (Jung et al., 2012, Kim et al., 2009, Kim et al., 2010, Kolekar et
362 al., 2013). Its ability to grow in alkaline conditions up to pH 12 has also been reported (Kim
363 et al., 2009, Tarhriz et al., 2012, Kim et al., 2010), and its ability to degrade a range of
364 substrates appears to have enhanced its ability to thrive within the CDP driven microcosm.
365 The ability to form biofilms and pellicles has been reported in *Alishewanella jeotgali* which
366 may indicate a pivotal role for the *Alishewanella* sp. in the formation and maintenance of the
367 bacterial aggregates within the microcosm (Jung et al., 2012) (Figure 5A).

368 A range of degradation rate constants for the various forms of ISA (alpha, beta and
369 xylo) were observed in the derived microcosms. The rate constant of beta ISA degradation
370 was similar to that reported by Rout et al (2015) at pH 11, whilst the rate constant of alpha
371 ISA degradation was greatly reduced, potentially due to the reduced role of key genera such
372 as *Alkaliphilus* (Rout et al., 2015). This is the first time that a microbial degradation rate
373 constant for xylo ISA has been published.

374 The loss of methanogens from the microcosm cannot be entirely attributed to
375 the pH, since a pH_{11.0} methanogenic microcosm has been successfully established using
376 sediments from the Buxton site (Rout et al., 2015). In that case a similar range of
377 methanogens were observed to that identified here associated with the cotton but with a
378 Eubacterial population dominated by *Alkaliphilus*. The lack of Clostridia species specifically
379 *Alkaliphilus* sp. within the microcosm formed from the colonised cotton appears to have
380 retarded the ability of the associated methanogenic population to become established. Tight
381 adherence to the cotton fibres and a possible differences in redox potential between the
382 internal biofilm environment and the enrichment media may have also contributed to the poor
383 transition of the methanogens and Clostridia species leading to an *Alishewanella* dominated
384 system (Sridhar and Eiteman, 1999, Stuart et al., 1999).

385 The presence of cotton fibres with the hyper alkaline analogue site at Harpur Hill
386 provided both a source of CDP to drive anoxic metabolism and a surface for microbial
387 colonisation. Subsequent sub culturing indicated that the cotton provided a surface for the
388 adherence of a narrow range of Clostridiaceae 2 species and promoted the development of a
389 floc based alkaliphilic population dominated by *Alishewanella* sp. able to degrade CDP up to
390 a pH of 11.0. Although methanogenic populations were detected on the cotton fibres, they
391 were unable to make the transition to floc based suspended growth.

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397 for emplacement.

398 **References**

- 399 Allard, S. & Ekberg, C. Complexing properties of α -Isosaccharinate: stability constants,
400 enthalpies and entropies of Th-complexation with uncertainty analysis. *Journal of*
401 *Solution Chemistry* 2006; **35**:1173-1186.
- 402 Almond, M., Shaw, P. B., Humphreys, P. N., Chadha, M. J., Niemelä, K. & Laws, A. P.
403 Behaviour of xyloisosaccharinic acid and xyloisosaccharino-1, 4-lactone in aqueous
404 solutions at varying pHs. *Carbohydrate Research* 2012; **363**:51-57.
- 405 Altschul, S. F., Madden, T. L., Schäffer, A. A., Zhang, J., Zhang, Z., Miller, W. & Lipman,
406 D. J. Gapped BLAST and PSI-BLAST: a new generation of protein database search
407 programs. *Nucleic Acids Research* 1997; **25**:3389-3402.
- 408 Andreesen, J. R., Gottschalk, G. & Schlegel, H. G. *Clostridium formicoaceticum* nov. spec.
409 isolation, description and distinction from *C. aceticum* and *C. thermoaceticum*.
410 *Archives of microbiology* 1970; **72**:154-174.
- 411 B.S.I. BS ISO 11271:2002 Soil quality — Determination of redox potential — Field method.
412 British Standards Institute, London, UK. 2002. BS ISO 11271:2002.
- 413 B.S.I. BS ISO 14853:2005 Plastics-Determination of the ultimate anaerobic biodegradation
414 of plastic materials in an aqueous system- Method by measurement of biogas
415 production. British Standards Institute, London, UK. 2005. BS ISO 14853:2005.
- 416 Bassil, N. M., Bryan, N. & Lloyd, J. R. Microbial degradation of isosaccharinic acid at high
417 pH. *The ISME Journal* 2014; **9**:310-320.
- 418 Bethke, C. M., Sanford, R. A., Kirk, M. F., Jin, Q. & Flynn, T. M. The thermodynamic ladder
419 in geomicrobiology. *American Journal of Science* 2011; **311**:183-210.
- 420 Burke, I. T., Mortimer, R. J., Palaniyandi, S., Whittleston, R. A., Lockwood, C. L., Ashley,
421 D. J. & Stewart, D. I. Biogeochemical reduction processes in a hyper-alkaline
422 leachate affected soil profile. *Geomicrobiology Journal* 2012; **29**:769-779.
- 423 Chapman, N. & Hooper, A. The disposal of radioactive wastes underground. *Proceedings of*
424 *the Geologists' Association*. 2012; **123**:46-63.
- 425 Cole, J. R., Wang, Q., Cardenas, E., Fish, J., Chai, B., Farris, R. J., Kulam-Syed-Mohideen,
426 A. S., Mcgarrell, D. M., Marsh, T., Garrity, G. M. & Tiedje, J. M. The Ribosomal
427 Database Project: improved alignments and new tools for rRNA analysis. *Nucleic*
428 *Acids Research* 2009; **37**:D141-D145.

- 429 Conrad, R., Ji, Y., Noll, M., Klose, M., Claus, P. & Enrich-Prast, A. Response of the
 430 methanogenic microbial communities in Amazonian oxbow lake sediments to
 431 desiccation stress. *Environmental Microbiology* 2014; **16**:1682-1694.
- 432 Das, T., Sehar, S., Koop, L., Wong, Y. K., Ahmed, S., Siddiqui, K. S. & Manefield, M.
 433 Influence of calcium in extracellular DNA mediated bacterial aggregation and biofilm
 434 formation. *PloS one* 2014; **9**:e91935.
- 435 Dell'anno, A. & Danovaro, R. Extracellular DNA plays a key role in deep-sea ecosystem
 436 functioning. *Science* 2005; **309**:2179.
- 437 Dominiak, D. M., Nielsen, J. L. & Nielsen, P. H. Extracellular DNA is abundant and
 438 important for microcolony strength in mixed microbial biofilms. *Environmental*
 439 *Microbiology* 2011; **13**:710-721.
- 440 Dridi, B., Fardeau, M.-L., Ollivier, B., Raoult, D. & Drancourt, M. *Methanomassiliicoccus*
 441 *luminyensis* gen. nov., sp. nov., a methanogenic archaeon isolated from human faeces.
 442 *International Journal of Systematic and Evolutionary Microbiology* 2012; **62**:1902-
 443 1907.
- 444 Eaton, A. D., Clesceri, L. S., Rice, E. W. & Greenberg, A. E. (eds.) 2005. *Standard Methods*
 445 *for the Examination of Water and Wastewater 21st Edition*, Washington, US:
 446 APHA/AWWA/WEF.
- 447 Edwards, U., Rogall, T., Blöcker, H., Emde, M. & Böttger, E. C. Isolation and direct
 448 complete nucleotide determination of entire genes. Characterization of a gene coding
 449 for 16S ribosomal RNA. *Nucleic Acids Research* 1989; **17**:7843-7853.
- 450 Fisher, E., Dawson, A. M., Polshyna, G., Lisak, J., Crable, B., Perera, E., Ranganathan, M.,
 451 Thangavelu, M., Basu, P. & Stolz, J. F. Transformation of inorganic and organic
 452 arsenic by *Alkaliphilus oremlandii* sp. nov. strain OhILAs. *Annals of the New York*
 453 *Academy of Sciences* 2008; **1125**:230-241.
- 454 Flemming, H. C. & Wingender, J. The biofilm matrix. *Nature Reviews Microbiology* 2010;
 455 **8**:623-33.
- 456 Gantner, S., Andersson, A. F., Alonso-Sáez, L. & Bertilsson, S. Novel primers for 16S
 457 rRNA-based archaeal community analyses in environmental samples. *Journal of*
 458 *Microbiological Methods* 2011; **84**:12-18.
- 459 Gloag, E. S., Turnbull, L., Huang, A., Vallotton, P., Wang, H., Nolan, L. M., Mililli, L.,
 460 Hunt, C., Lu, J. & Osvath, S. R. Self-organization of bacterial biofilms is facilitated
 461 by extracellular DNA. *Proceedings of the National Academy of Sciences* 2013;
 462 **110**:11541-11546.
- 463 Gorlenko, V., Tsapin, A., Namsaraev, Z., Teal, T., Tourova, T., Engler, D., Mielke, R. &
 464 Nealsen, K. *Anaerobranca californiensis* sp. nov., an anaerobic, alkalithermophilic,
 465 fermentative bacterium isolated from a hot spring on Mono Lake. *International*
 466 *Journal of Systematic and Evolutionary Microbiology* 2004; **54**:739-743.
- 467 Grant, W. D., Holtom, G. J., O'Kelly, N., Malpass, J., Rosevear, A., Watkiss, P. &
 468 Widdowson, D. 2002. Microbial degradation of cellulose-derived complexants under
 469 repository conditions. AEA Technology and University of Leicester for UK Nirex
 470 Ltd. AEAT/ERRA-0301.
- 471 Greenfield, B. F., Moreton, A. D., Spindler, M. W., Williams, S. J. & Woodwark, D. R. The
 472 effects of the degradation of organic materials in the near field of a radioactive waste
 473 repository. *MRS Online Proceedings Library* 1991; **257**.

474 Humphreys, P., Laws, A. & Dawson, J. A review of cellulose degradation and the fate of
475 degradation products under repository conditions. SERCO / TAS / 002274 / 001. Serco
476 contractors report for the Nuclear Decommissioning Authority, UK 2010a.

477 Humphreys, P. N., West, J. M. & Metcalfe, R. 2010b. Microbial effects on repository
478 performance. Quintessa contractors report for the Nuclear Decommissioning
479 Authority, UK. QRS-1378Q-1.

480 Jahn, A. & Nielsen, P. H. Cell biomass and exopolymer composition in sewer biofilms.
481 *Water Science and Technology* 1998; **37**:17-24.

482 Jones, B. E., Grant, W. D., Collins, N. C. & Mwatha, W. E. 1994. Alkaliphiles: diversity and
483 identification. *Bacterial Diversity and Systematics*. Springer.

484 Jung, J., Chun, J. & Park, W. Genome sequence of extracellular-protease-producing
485 *Alishewanella jeotgali* isolated from traditional Korean fermented seafood. *Journal of*
486 *Bacteriology* 2012; **194**:2097.

487 Kim, M.-S., Jo, S. K., Roh, S. W. & Bae, J.-W. *Alishewanella agri* sp. nov., isolated from
488 landfill soil. *International Journal of Systematic and Evolutionary Microbiology*
489 2010; **60**:2199-2203.

490 Kim, M. S., Roh, S. W., Nam, Y. D., Chang, H. W., Kim, K. H., Jung, M. J., Choi, J. H.,
491 Park, E. J. & Bae, J. W. *Alishewanella jeotgali* sp. nov., isolated from traditional
492 fermented food, and emended description of the genus *Alishewanella*. *International*
493 *Journal of Systematic and Evolutionary Microbiology* 2009; **59**:2313-6.

494 Knill, C. J. & Kennedy, J. F. Degradation of cellulose under alkaline conditions.
495 *Carbohydrate Polymers* 2003; **51**:281-300.

496 Kolekar, Y. M., Konde, P. D., Markad, V. L., Kulkarni, S. V., Chaudhari, A. U. & Kodam, K.
497 M. Effective bioremoval and detoxification of textile dye mixture by *Alishewanella*
498 sp. KMK6. *Applied Microbiology and Biotechnology* 2013; **97**:881-889.

499 Kotelnikova, S., Macario, A. J. & Pedersen, K. *Methanobacterium subterraneum* sp. nov., a
500 new alkaliphilic, eurythermic and halotolerant methanogen isolated from deep granitic
501 groundwater. *International Journal of Systematic Bacteriology* 1998; **48**:357-367.

502 Libert, M., Bildstein, O., Esnault, L., Jullien, M. & Sellier, R. Molecular hydrogen: An
503 abundant energy source for bacterial activity in nuclear waste repositories. *Physics*
504 *and Chemistry of the Earth, Parts A/B/C* 2011; **36**:1616-1623.

505 Maestrojuan, G. M., Boone, J. E., Mah, R. A., Menaia, J. A., Sachs, M. S. & Boone, D. R.
506 Taxonomy and halotolerance of mesophilic *Methanosarcina* strains, assignment of
507 strains to species, and synonymy of *Methanosarcina mazei* and *Methanosarcina*
508 *frisla*. *International Journal of Systematic Bacteriology* 1992; **42**:561-567.

509 McDonald, J. E., Houghton, J. N., Rooks, D. J., Allison, H. E. & McCarthy, A. J. The
510 microbial ecology of anaerobic cellulose degradation in municipal waste landfill sites:
511 evidence of a role for fibrobacters. *Environmental Microbiology* 2012; **14**:1077-1087.

512 Motellier, S., Richet, C. & Merel, P. Analysis of cellulose degradation products by capillary
513 electrophoresis. *Journal of Chromatography A* 1998; **804**:363-370.

514 N.D.A. Geological Disposal: An introduction to the generic disposal system safety case.
515 Nuclear Decommissioning Authority. Harwell, Didcot, Oxfordshire, UK. 2010a.

516 N.D.A. Near-field Evolution Status Report Nuclear Decommissioning Authority. Harwell,
517 Didcot, Oxfordshire, UK 2010b. NDA/RWMD/033.

518 N.D.A. Radioactive Wastes in the UK: A Summary of the 2013 Inventory. Nuclear
519 Decommissioning Authority. Harwell, Didcot, Oxfordshire, UK. 2013.

520 Ordoñez, O. F., Flores, M. R., Dib, J. R., Paz, A. & Farías, M. E. Extremophile culture
521 collection from Andean lakes: extreme pristine environments that host a wide
522 diversity of microorganisms with tolerance to UV radiation. *Microbial Ecology* 2009;
523 **58**:461-473.

- 524 Pikuta, E. V., Itoh, T., Krader, P., Tang, J., Whitman, W. B. & Hoover, R. B. *Anaerovirgula*
525 *multivorans* gen. nov., sp. nov., a novel spore-forming, alkaliphilic anaerobe isolated
526 from Owens Lake, California, USA. *International Journal of Systematic and*
527 *Evolutionary Microbiology* 2006; **56**:2623-9.
- 528 Randall, M., Rigby, B., Thomson, O. & Trivedi, D. Assessment of the effects of cellulose
529 degradation products on the behaviour of europium and thorium. Report prepared on
530 behalf of NDA by National Nuclear Laboratory. 2013. NNL (12) 12239 Part A, Issue
531 4.
- 532 Rittmann, B. E. & Mccarty, P. L. *Environmental biotechnology*, McGraw Hill New
533 York.2001
- 534 Rizoulis, A., Steele, H., Morris, K. & Lloyd, J. The potential impact of anaerobic microbial
535 metabolism during the geological disposal of intermediate-level waste. *Mineralogical*
536 *Magazine* 2012; **76**:3261-3270.
- 537 Rodrigues, D. F., Goris, J., Vishnivetskaya, T., Gilichinsky, D., Thomashow, M. F. & Tiedje,
538 J. M. Characterization of *Exiguobacterium* isolates from the Siberian permafrost.
539 Description of *Exiguobacterium sibiricum* sp. nov. *Extremophiles* 2006; **10**:285-294.
- 540 Rout, S. P., Charles, C. J., Garratt, E. J., Laws, A. P., Gunn, J. & Humphreys, P. N. Evidence
541 of the generation of isosaccharinic acids and their subsequent degradation by local
542 microbial consortia within hyper-alkaline contaminated soils, with relevance to
543 intermediate level radioactive waste disposal. *PloS one* 2015; **10**:e0119164.
- 544 Rout, S. P., Radford, J., Laws, A. P., Sweeney, F., Elmekawy, A., Gillie, L. J. & Humphreys,
545 P. N. Biodegradation of the alkaline cellulose degradation products generated during
546 radioactive waste disposal. *PloS one* 2014; **9**:e107433.
- 547 Schloss, P. D., Westcott, S. L., Ryabin, T., Hall, J. R., Hartmann, M., Hollister, E. B.,
548 Lesniewski, R. A., Oakley, B. B., Parks, D. H. & Robinson, C. J. Introducing mothur:
549 open-source, platform-independent, community-supported software for describing and
550 comparing microbial communities. *Applied and Environmental Microbiology* 2009;
551 **75**:7537-7541.
- 552 Shaw, P. B., Robinson, G. F., Rice, C. R., Humphreys, P. N. & Laws, A. P. A robust method
553 for the synthesis and isolation of β -gluco-isosaccharinic acid ((2R,4S)-2,4,5-
554 trihydroxy-2-(hydroxymethyl)pentanoic acid) from cellulose and measurement of its
555 aqueous pKa. *Carbohydrate Research* 2012; **349**:6-11.
- 556 Sridhar, J. & Eiteman, M. A. Influence of redox potential on product distribution in
557 *Clostridium thermosuccinogenes*. *Applied Biochemistry and Biotechnology* 1999;
558 **82**:91-101.
- 559 Stuart, S. L., Woods, S. L., Lemmon, T. L. & Ingle, J. D. The effect of redox potential
560 changes on reductive dechlorination of pentachlorophenol and the degradation of
561 acetate by a mixed, methanogenic culture. *Biotechnology and Bioengineering* 1999;
562 **63**:69-78.
- 563 Summons, R. E., Schubotz, F., Hays, L. E., Meyer-Dombard, D., Gillespie, A. & Shock, E.
564 Stable isotope labeling confirms mixotrophic nature of streamer biofilm communities
565 at alkaline hot springs. *Name: Frontiers in Microbiology* 2015; **6**:42.
- 566 Takai, K., Moser, D. P., Onstott, T. C., Spoelstra, N., Pfiffner, S. M., Dohnalkova, A. &
567 Fredrickson, J. K. *Alkaliphilus transvaalensis* gen. nov., sp. nov., an extremely
568 alkaliphilic bacterium isolated from a deep South African gold mine. *International*
569 *Journal of Systematic and Evolutionary Microbiology* 2001; **51**:1245-56.
- 570 Tarhriz, V., Nematzadeh, G., Vahed, S. Z., Hejazi, M. A. & Hejazi, M. S. *Alishewanella*
571 *tabrizica* sp. nov., isolated from Qurugöl Lake. *International Journal of Systematic*
572 *and Evolutionary Microbiology* 2012; **62**:1986-1991.

- 573 Urbietta, M. S., González-Toril, E., Bazán, Á. A., Giaveno, M. A. & Donati, E. Comparison
574 of the microbial communities of hot springs waters and the microbial biofilms in the
575 acidic geothermal area of Copahue (Neuquén, Argentina). *Extremophiles* 2015:1-14.
- 576 Van Loon, L. R. & Glaus, M. A. Review of the kinetics of alkaline degradation of cellulose
577 in view of its relevance for safety assessment of radioactive waste repositories.
578 *Journal of Environmental Polymer Degradation* 1997; **5**:97-109.
- 579 Warwick, P., Evans, N., Hall, T. & Vines, S. 2003. Complexation of Ni(II) by α -
580 isosaccharinic acid and gluconic acid from pH 7 to pH 13. *Radiochimica*
581 *Acta/International journal for chemical aspects of nuclear science and technology*.
- 582 Williamson, A. J., Morris, K., Shaw, S., Byrne, J. M., Boothman, C. & Lloyd, J. R. Microbial
583 reduction of Fe(III) under alkaline conditions relevant to geological disposal. *Applied*
584 *Environmental Microbiology* 2013; **79**:3320-3326.
- 585 Worakit, S., Boone, D. R., Mah, R. A., Abdel-Samie, M.-E. & El-Halwagi, M.
586 *Methanobacterium alcaliphilum* sp. nov., an H₂-utilizing methanogen that grows at
587 high pH values. *International Journal of Systematic Bacteriology* 1986; **36**:380-382.
- 588 Wu, W.-M., Jain, M. K., De Macario, E. C., Thiele, J. H. & Zeikus, J. G. Microbial
589 composition and characterization of prevalent methanogens and acetogens isolated
590 from syntrophic methanogenic granules. *Applied Microbiology and Biotechnology*
591 1992; **38**:282-290.
- 592

592

593 Table 1: Analysis of porewater, sediment and cotton retrieved from sample borehole.

594

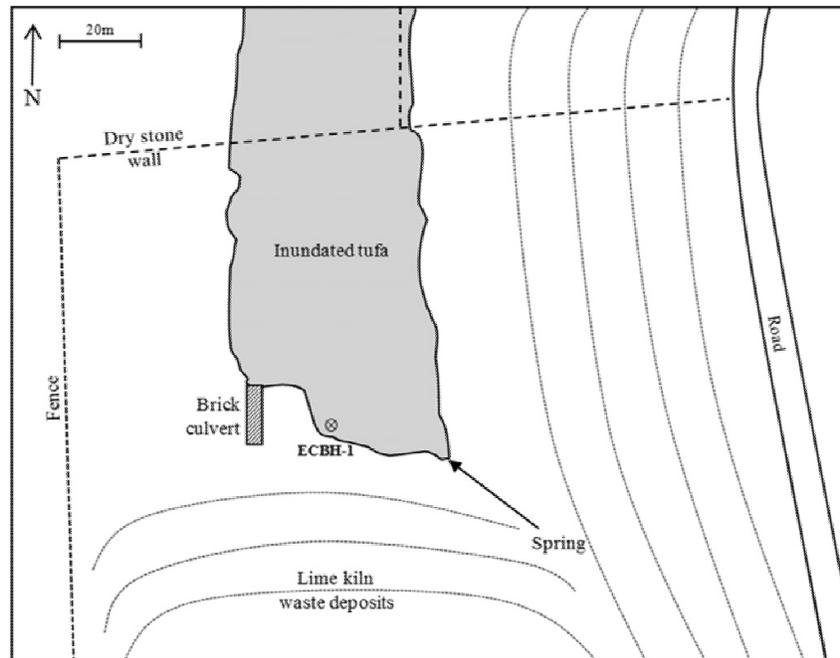
Source	pH	eH	Acetate	α -ISA	β -ISA
Porewater (mg L ⁻¹)	11.92	-66.00	208.90	7.64	6.82
Sediment (mg (g dry wt) ⁻¹)					
1)	11.50	-77.00	127.24	1.01	0.54
Cotton (mg (g dry wt) ⁻¹)	N/S	N/S	141.16	2.34	0.85

595 *N/S-Not sampled

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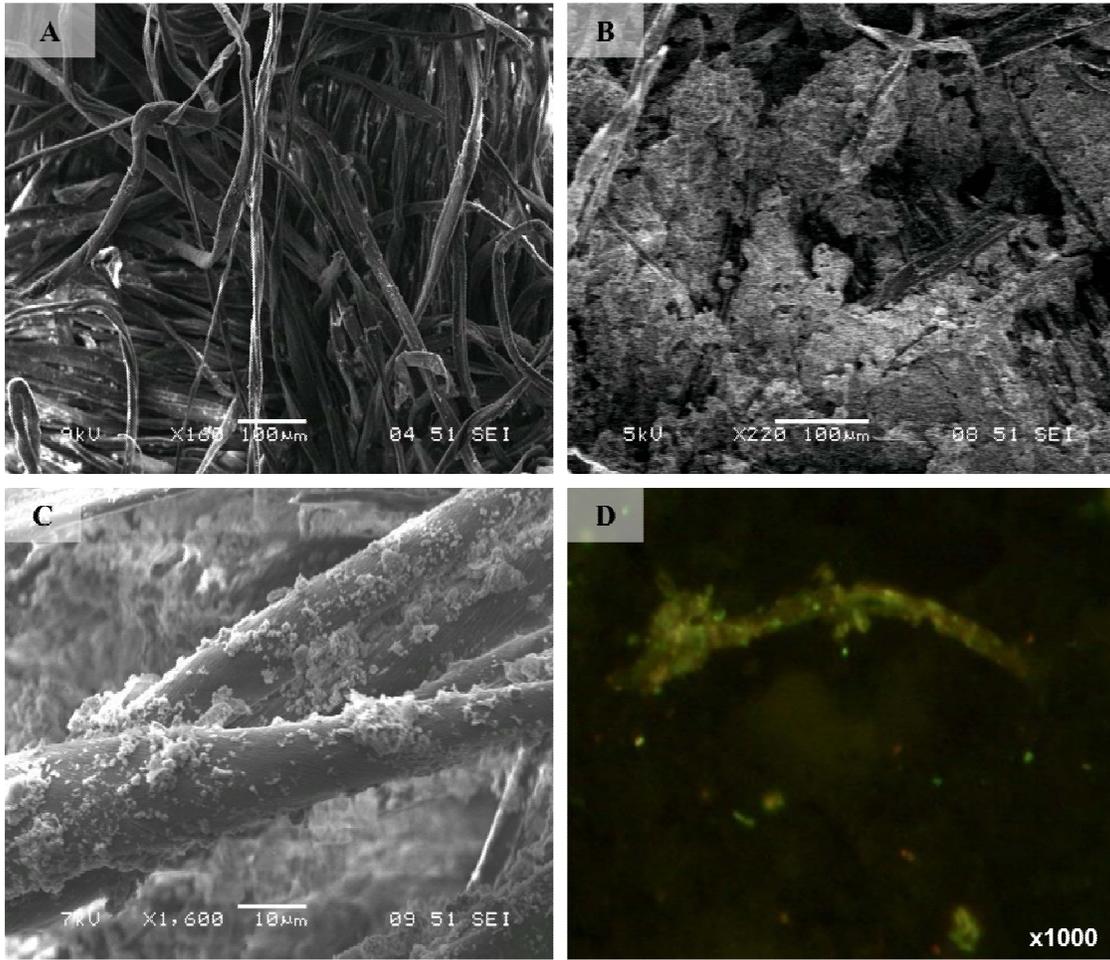


598

599 Figure 1: Overview of hyperalkaline contaminated site and position of emplaced
600 cotton within bore hole 1 (ECBH-1)

601

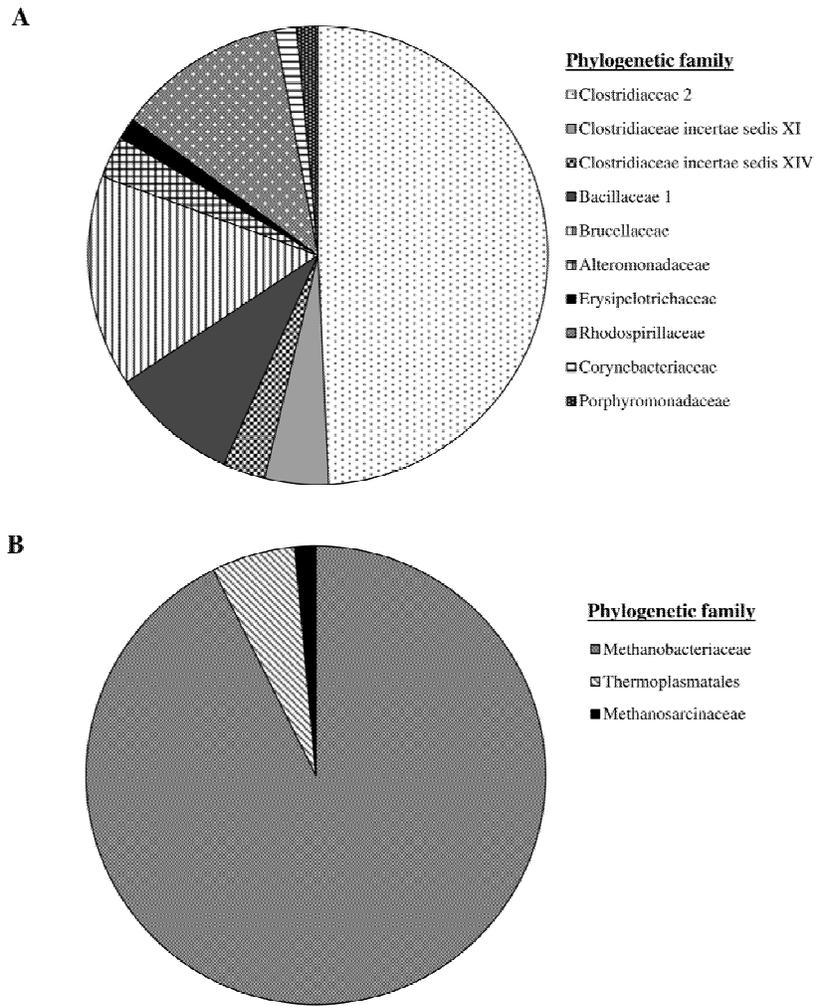
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603 Figure 2: Microscopy investigation of the cellulose cotton. [A] Sterile cotton. [B]
604 Cellulose cotton from the borehole showing biofilm formation. [C] Close up of
605 individual fibre showing individual cells, EPS aggregates and mineral precipitate. [D]
606 Live/dead image of individual cotton fibre.

607



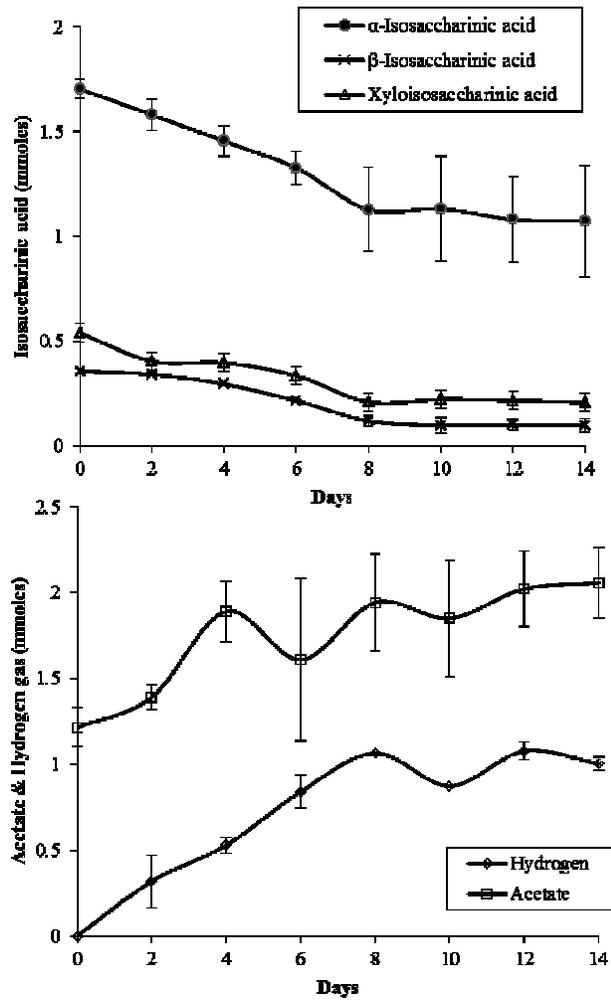
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610 Figure 3: 16S rRNA gene clone libraries of the colonised cotton. [A] Eubacterial
 611 (n=67). [B] Archaeal (n=68). Phylogenetic families were assigned to clones through a
 612 MegaBLAST database search.

613

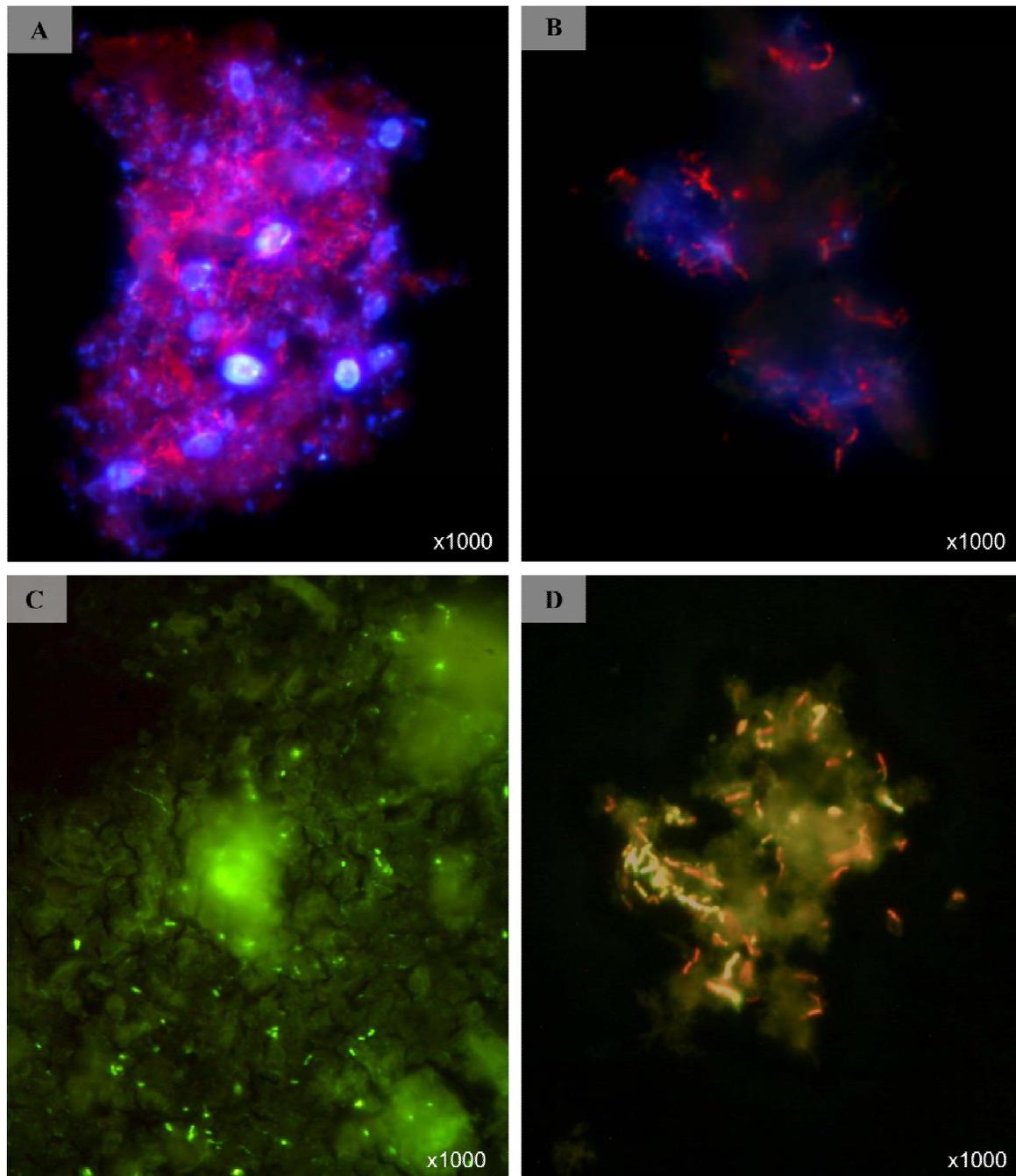
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615 Figure 4: Chemistry of the CDP driven pH 11 microcosm over two waste/feed cycles
616 using colonised cellulose cotton as an inoculation source. [A] Alpha, beta and xylo
617 isosaccharinic acid degradation profile. [B] Hydrogen and acetate production profile.

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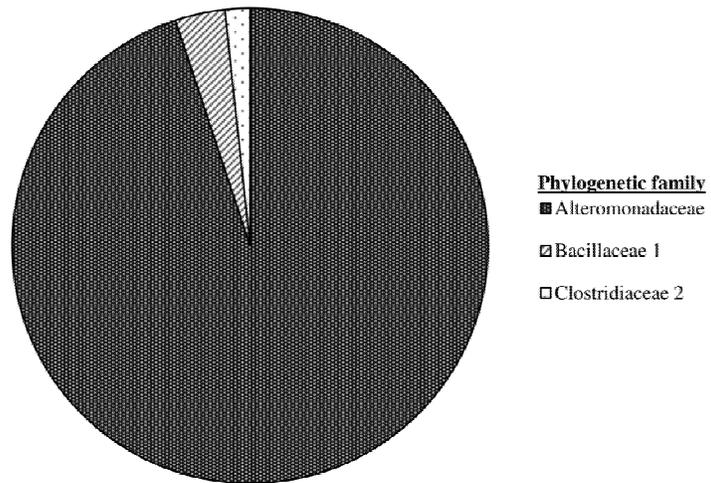
620 Figure 5: Microscopy investigation into the morphology of the pH 11 microcosm. [A]
 621 Ethidium bromide and Calcofluor white stain of bacterial flocculate showing individual
 622 cells and eDNA (red) and extracellular polysaccharides (blue). [B] DNase digest of
 623 flocculates stained with ethidium bromide and Calcofluor white. [C] FITC stain of

624 bacterial flocculate showing areas containing protein (green). [D] Live/dead image of

625 bacterial flocculate.

626

626



627

628 Figure 6: Eubacterial (n=59) 16s rRNA gene clone library of pH 11 CDP driven
629 microcosm.

630

631